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1	Removal of Microcystis aeruginosa by UV-activated persulfate: Performance and
2	characteristics
3	
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13	Abstract
14	Cyanobacteria blooms in source waters have become a worldwide issue for drinking
15	water production. UV-activated persulfate (UV/PS) technology was firstly applied to
16	remove cultivated Microcystis aeruginosa in bench scale. The presence of persulfate
17	significantly enhanced both cytoclasis and algal organic matter mineralization
18	compared with UV-C inactivation alone. Around 98.2% of algal cells were removed
19	after UV/PS process treatment for 2 h at a dosage of PS being 1500 mg/L
20	(approximately 6 mM). Both sulfate and hydroxyl radicals were proven to contribute

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21	to the removal of algae and the loss of cell integrity. The cultivated Microcystis
22	aeruginosa in death growth phase were found to be more vulnerable to UV/PS
23	treatment than those growing in log phase, thus a significant lower dosage of PS is
24	needed to achieve the desired removal efficiency. This study suggested a novel
25	application of UV/PS process in the removal of algae in source waters due to the high
26	degradation efficiency of both algal cells and their derived organic matter.

Keywords: Microcystis aeruginosa; Ultraviolet; Persulfate; Cell integrity; Algal
organic matter

32 **1. Introduction**

33 In recent decades, harmful algal blooms have frequently occurred in source waters 34 such as lakes, rivers and reservoirs associated with eutrophication throughout the 35 world [1-4]. Cyanobacteria, a prominent and ubiquitous issue, has attracted 36 worldwide attention among the harmful algal blooms. Along with excessive algal cells, 37 the algal organic matter (AOM) including extracellular organic matter (EOM) and 38 intracellular organic matter (IOM) generated via metabolic excretion always cause 39 serious water problems [5]. The AOM produced by some specific Cyanobacteria 40 genera including anabaena, microcystis, planktothrix, etc. has been proven to comprise taste- and odor- substances, such as 2-methylisoborneol and geosmin [6], 41 42 and a wide range of toxic intracellular metabolites being suggested to cause both acute 43 and chronic effects on hepatocyte and central nervous system of aquatic organisms 44 and biomagnify [7-10]. Moreover, algal cells and AOM have been reported to be 45 important precursors of vast disinfection by-products (DBPs) including traditional 46 trihalomethanes (THMs), haloacetic acids (HAAs) and emerging nitrogenous DBPs 47 (N-DBPs) containing high genotoxicity and carcinogenicity [11-13].

However, the conventional drinking water treatment plant only shows limited removal efficiency on algal cells due to electrostatic repulsion, surface hydrophilicity and steric effects [14-16]. The residual cells after sedimentation could adhere to the filter material surface subsequently causing filter clogging, penetrate into the water supply pipe network, and finally impair the drinking water quality [17]. It should be 53 noted that the dissolved AOM generated via metabolic excretion are even more 54 different than algal cells be removed by using traditional to 55 coagulation-sedimentation-filtration process [18], which may adversely affect 56 conventional water production via inhibition of coagulation [19, 20].

57 Activated persulfate (PS) oxidation has been studied as an alternative conventional 58 advanced oxidation process (AOP) in water treatment [21-23]. The AOP using PS is mainly achieved by the formation of reactive sulfate radical (SO₄^{-,}, E^0 =2.65–3.1 V) 59 through the decomposition of PS by heat, transit metals, light, microwave or 60 61 ultrasound [24-26]. Similar to hydroxyl radical (HO[•], E^0 =1.8–2.7 V), electron-transfer is expected to be a vital reaction when SO4⁻ is used to degrade organic pollutants [27]. 62 63 However, SO4⁻⁻ is more selective in comparison with hydroxyl radical in general, thus 64 might be more effective in the degradation of some organic pollutants in the presence 65 of radical scavengers [27]. Particularly, PS activated by zerovalent iron was recently used for disinfection of ballast water and achieved a result that the species of marine 66 67 phytoplankton could be inactivated by such a process without generating harmful 68 byproducts [28]. However, to our knowledge, activated PS technology has not been 69 utilized for algae removal in source waters.

Among the known PS-activation processes, ultraviolet (UV) radiation at 254 nm can activate the PS to generate SO_4^{-} through Eq. 1 with a quantum yield of 0.7 mol E/s [29], and the other main chemical interactions subsequently take place were summarized in Eqs. 2 ~ 5 [30].

74
$$S_2 O_8^{2-} \xrightarrow{UV} 2SO_4^{\bullet-} \qquad \phi = 0.7$$
 (1)

75
$$SO_4^{\bullet-} + OH^- \longrightarrow SO_4^{2-} + HO^{\bullet} \qquad k = 6.5 \times 10^7 \,\mathrm{M}^{-1}\mathrm{s}^{-1}$$
 (2)

76
$$\operatorname{SO}_{4}^{\bullet-} + \operatorname{H}_{2}\operatorname{O} \longrightarrow \operatorname{HSO}_{4}^{2-} + \operatorname{HO}^{\bullet} \qquad k = 8.3 \text{ M}^{-1} \text{s}^{-1}$$
(3)

77
$$OH^{\bullet-} + S_2 O_8^{2-} \longrightarrow S_2 O_8^{\bullet-} + OH^- \qquad k = 1.4 \times 10^7 \text{ M}^{-1} \text{s}^{-1}$$
(4)

78
$$OH^{\bullet-} + SO_4^{\bullet-} \longrightarrow HSO_5^{-} \qquad k = 1 \times 10^{10} M^{-1} s^{-1}$$
 (5)

Furthermore, UV process has been applied to suppress algae growth in many cases accompanied by chlorophyll bleaching and inhibition of metabolic activity [31-35]. Therefore, the UV radiation can be regarded as one of the feasible techniques to activate the PS for drinking water treatment applications.

In the present study, the performance of UV-activated PS (UV/PS) process on algae removal has been investigated using *Microcystis aeruginosa* (*M. aeruginosa*), a most abundant and common occurring cyanobacteria specie [36]. Additionally, variations in the characteristics of AOM during the UV/PS treatment were identified to further recognize the plausible by-products within the oxidation process.

88

89 2. Materials and methods

90 2.1. Materials

91 The *M. aeruginosa* (No. FACHB-909) was purchased from the Institute of 92 Hydrobiology, Chinese Academy of Sciences, and laboratorial cultivated in a light 93 growth incubator (Guohua Electric Co., Ltd., China). The *M. aeruginosa* was 94 incubated under the specific growth conditions (25 ± 1 °C with a light-dark cycle of

95	12 h: 12 h) in BG-11 media [37]. The algae solutions in log phase were harvested and
96	diluted using ultrapure water (18.5 M Ω cm) produced from a water purifier (PCDX-J,
97	Pincheng Co. Ltd, China). Then a final cell density of 1×10^6 cell/mL was achieved
98	for the experiments, which mimics to the practical density in harmful algal blooms
99	[38].
100	All chemicals used in the experiments were of analytical reagent grade at least.

101 Sodium persulfate (Na₂S₂O₈), sodium hydroxide, hydrochloric acid, methanol (MeOH)

102 and tert-butyl-alcohol (TBA) were obtained from Sinopharm Chemical Reagent Co.,

Ltd., China. Powdered Na₂S₂O₈ was added to the prepared *M. aeruginosa* solutions to
achieve designated concentrations of PS each time. Solution pH was subsequently
adjusted to be 7.0 which is around the pH of natural water by using HCl and NaOH at
a concentration of 0.1 M.

107

108 2.2. Experimental reactor and procedures

Algae removal experiments were carried out in a cylindrical pyrex reactor (600 mL effective volume) with a low-pressure UV lamp (254 nm, 23W, GPH 436T5L/4, Philips Electronics Ltd., The Netherlands) (Fig. S1). An immersion well made of high purity quartz was placed inside the glass reactor. The UV lamp was fixed inside the immersion well. Cooling water was pumped through the thin annular zone of the immersion well to prevent overheating of the reaction solutions. In order to achieve a stabilized radiation intensity (measured to be 1.25 mW/cm² in average by the reported 116 method [39]), the lamp was always switched on for 15 min before being placed into the reactor. A magnetic stirring apparatus at a speed of 200 rpm was used to 117 118 homogenize the solutions throughout the experiments. Samples were collected via the 119 sampling port at specific time intervals. To evaluate the reaction mechanisms in the 120 UV/PS system, MeOH and TBA were added as scavengers for hydroxyl and sulfate 121 radicals. Each batch of experiment was carried out in triplicate. As the formed SO4⁻⁻ could be scavenged by high concentrations of S₂O₈²⁻ and Cl⁻ from BG-11 media and 122 123 HCl solution (section 3.2), the presence of anions including NO₃⁻ and HCO₃⁻ at 124 concentrations as usual in natural waters was expected to play little role in the removal of algal cells (Fig. S2). Thus, the impacts of co-existing anions would not be 125 126 further discussed in the following sections.

127

128 2.3. Analytical methods

129 The cell concentration of *M. aeruginosa* was measured using an UV-visible 130 spectrophotometer (U-3100, Hitachi, Japan) at a wavelength of 681 nm since the 131 optical density at 681 nm (OD₆₈₁) is linearly correlated with counted cell number by 132 microscope within the experimental range [40, 41]. Thus, the removal efficiency of 133 algal cells (ρ , %) can be calculated using Eq. 6.

134
$$\rho = \frac{(OD_{681_0} - OD_{681_1})}{OD_{681_0}} \times 100\%$$
(6)

135 where OD_{681_0} and OD_{681_1} were the optical density values at 0 min and t min.

136 Chlorophyll-a (Chl-a) was extracted using acetone solution and then measured

using the spectrophotometer at wavelengths of 663 nm, 645 nm and 630 nm accordingto the reported method [42].

139 The algal cell integrity before and after treatment was monitored by a flow 140 cytometer (Accuri C6, BD Biosciences, USA) equipped with an argon laser emitting 141 (wavelength fixed at 488 nm) for fluorescence measurement. Algal cells were stained 142 using SYTOX green nucleic acid stain (Invitrogen, Life Technologies, USA) [43]. 143 Fluorescent filters and detectors were equipped to collect green fluorescence in channel FL1 (530 nm) and red fluorescence in channel FL3 (630 nm), which 144 145 represents damaged and integrated cells, respectively, since SYTOX could penetrate 146 damaged cells and stain the nucleic acid to emit green fluorescence.

147 Extracellular AOM was extracted from reaction solution by centrifuging the cell 148 suspension at 8000 rmp for 15 min and subsequently filtering the supernatant through 149 0.45 µm cellulose acetate membranes [5]. The extracellular AOM was characterized 150 on fluorescence spectroscopy, UV-Vis spectrum scan, and total organic carbon (TOC). 151 A fluorescence spectrophotometer (F-4600, Hitachi, Japan) was used to measure the 152 fluorescence excitation-emission matrix (EEM) spectroscopy of AOM. Excitation wavelengths (Ex) were scanned from 200 to 450 nm with 5 nm intervals and emission 153 154 wavelengths (Em) from 280 to 550 nm with 2 nm intervals. The scanning speed was 155 set at 1200 nm/min. Background signals were minimized by subtracting the signals of 156 the blank (i.e. ultrapure water). The EEM data were analyzed by MATLAB 2010b (The MathWorks, Inc., USA). The UV-Vis spectrum scan of AOM ranging from 200 157

to 700 nm was measured using the U-3100 spectrophotometer. The concentration of
dissolved organic carbon (DOC) in the prepared sample was measured using a
TOC/TN analyzer (C/N 2100, Analytic Jena, Germany).

161 The persulfate concentration was measured by universal iodometric titration 162 method [44]. The procedure was conducted by mixing 2 mL samples and 10 mL KI 163 stock solution (10% mass fraction) in 40 mL ultrapure water. Then the resulting 164 solutions were equilibrated for 3 h in dark, and subsequently titrated by calibrated Na₂S₂O₃ solution (0.1 M). A turbidimeter (Ruixin WGZ-2, China) was used to 165 166 measure the turbidity of algal solution before and after treatment. The M. aeruginosa cells in suspension for morphology observation were firstly centrifuged at 6000 rpm 167 to collect the precipitate, and then dried by a vacuum freeze dryer (model FD-1A-50, 168 169 Shanghai Boyikang Instrument Co., Ltd., China). Thereafter, the dried algae samples 170 were sputter coated with gold by the sputter coater, and then photographed using a 171 scanning electron microscopy (SEM) (Sirion 200, FEI, USA) at 10 kV.

172

173 **3. Results and discussion**

174 3.1. Comparison of UV, PS and UV/PS processes on algae removal

Fig. 1 shows the removal of *M. aeruginosa* based on the changes of OD₆₈₁ (Fig. 1a)
and Chl-a (Fig. 1b) by using three different oxidation processes, UV irradiation, PS
oxidation and combined UV/PS oxidation, at 25 °C with the initial reaction pH being
7.0. Negligible decrease of OD₆₈₁ and Chl-a was observed using PS oxidation alone at

179 a concentration of 1500 mg/L (about 6 mM) over a period of 120 min, indicating little to no oxidation of algal cells by PS directly. In contrast, UV irradiation could degrade 180 181 algal cells to some extent with OD₆₈₁ and Chl-a being removed by 21.8% and 87.9%, 182 respectively, after 120 min of radiation. The results corroborate that UV at 254 nm can 183 effectively destroy the photosynthesis system of *M. aeruginosa*, but hardly cause 184 large-scale cell cytolysis, which was also reported in other publications [32, 45]. 185 UV/PS process showed the highest levels of removal efficiency for both OD₆₈₁ 186 (98.2%) and Chl-a (99.8%) among the three oxidation processes.

187 As PS oxidation and UV irradiation had limited degradation efficiency on algal cells, it could be expected that the reactive radicals, i.e. SO4⁻⁻ and HO[•], produced in 188 189 UV/PS system were responsible for the observed degradation of algal cells. The 190 conclusion was further proved by the results shown in Fig. S3 that the presence of 191 MeOH and TBA inhibited the removal of OD₆₈₁ significantly when using UV/PS 192 process. MeOH is usually considered as an effective quencher for both the SO₄⁻⁻ and 193 HO[•] [46]. However, TBA is an effective quencher for HO[•], but not for SO₄^{•-} [29]. 194 Thus, they could be used to distinguish SO4⁻⁻ and HO⁻ based on the difference of the 195 degradation rate of target compounds. Fig. S3 shows that co-existence of either 196 MeOH or TBA inhibited the removal of OD₆₈₁, while the inhibiting ability of MeOH 197 is stronger than that of TBA. It proved that both SO₄⁻⁻ and HO[•] contributed to the 198 oxidation of algal cells in the experiment. It is known that activation of PS by UV only generate SO4⁻⁻ (Eq. 1) without the production of HO⁻ [29]. However, HO⁻ was 199

proposed to be generated when SO_4^{-} reacts with OH^- at a rate constant of 6.5×10^7 $M^{-1}S^{-1}$ through Eq. 4 [30], or with water at a rate constant of 8.3 M⁻¹S⁻¹ through Eq. 3 [47]. Besides, the presence of chloride could convert SO_4^{-} to HO[•] through reactions shown in Eqs. 7 ~ 10 [48], especially when the reaction pH was higher than 5.0, which make sure the presence of HO[•] in the experiment [49].

$$SO_4^{\bullet-} + Cl^- \overleftrightarrow{SO_4^{2-}} + Cl^{\bullet}$$
(7)

$$206 Cl^{\bullet} + OH^{-} \longrightarrow HO^{\bullet} + Cl^{-} (8)$$

207
$$\operatorname{Cl}^{\bullet} + \operatorname{H}_2 O \longrightarrow \operatorname{HOCl}^{\bullet-} + \operatorname{H}^+$$
 (9)

$$HOCl^{\bullet-} \xrightarrow{} HO^{\bullet} + Cl^{-}$$
(10)

209 3.2. Effect of initial PS dose on algae removal

210 The effect of initial PS dose ranging from 0 mg/L to 1500 mg/L on the removal of M. 211 aeruginosa by focusing on the changes of OD₆₈₁ and Chl-a was investigated (Fig. 2). 212 The removal rates of OD₆₈₁ and Chl-a increased with the increase of initial PS dosage. 213 When the dosage of PS increased from 0 mg/L to 1500 mg/L (about 6 mM), the 214 removal rates of OD₆₈₁ and Chl-a increased from 21.8% and 87.9% to 98.2% and 215 99.8%, respectively, after treatment for 120 min. It should be noted that, the rapid 216 decrease of Chl-a was observed in the initial 60 min treatment (including the only UV 217 irradiation case), but not for the removal of OD₆₈₁. The result suggested that the photosynthetic system of *M. aeruginosa* was damaged immediately through 218 219 synergetic irradiation and oxidation in UV/PS system, possibly due to the rapid 220 destruction in gene expression of both *psbA* (for D1) and *cpc* (for phycocyanin) [50].

The variation of pHs over time under different initial PS doses was displayed in Fig. S4. Negligible change of pHs was observed under UV irradiation alone over a period of 120 min. However, solution pHs gradually reduced in UV/PS system during the oxidation, which was enhanced with the increase of PS doses. The results can be explained by the production of sulfate acid through $S_2O_8^{-1}$ oxidation of water (Eqs. 11 and 12) [30, 51]:

227
$$SO_4^{\bullet-} + S_2O_8^{2-} \longrightarrow S_2O_8^{\bullet-} + SO_4^{2-}$$
(11)

228
$$S_2O_8^{\bullet-} + \frac{3}{2}H_2O \longrightarrow 2SO_4^{2-} + \frac{3}{2}O_2 + 3H^+$$
 (12)

Fig. S5 shows that the dissolved oxygen (DO) in the experimental solution significantly increased in the initial 30 min when using UV/PS process to treat algae, supporting the speculation on oxidation of water by S_2O_8 . Then the gradually reduction of DO in the subsequent 90 min might due to the release into air, reacting with HO[•] adducts to produce peroxyl transients [52], and being converted to oxygen-centered radicals which could be consumed by algae and its derived organic matter [53].

If the dosed persulfate was completely decomposed, formed sulfate at concentrations of 800 mg/L and 1200 mg/L would be achieved when the dosages of persulfate were 1000 mg/L and 1500 mg/L, respectively. A sulfate concentration that is higher than 600 mg/L in drinking water might cause taste alteration and diarrhea [54]. However, it should be noted that the concentrations of *M. aeruginosa* in most source waters were always lower than 10^6 cells/mL. When the initial cell density was reduced to 2.3×10^5 cells/mL, the residual OD₆₈₁ and turbidity decreased by 57.9% and 62.3%, respectively, after UV/PS process treatment for 120 min with an initial PS dosage of 200 mg/L (Fig. S6).

- 245
- 246 3.3. Effect of growth phase on algae removal

247 The growth of *M. aeruginosa* can be modeled with four different phases: lag phase, 248 log phase, stationary phase, and death phase. The removal of algae in their log phase 249 and death phase were investigated due to the large variation of cell population and 250 dissolved organic substances [55]. The effect of growth phase on the removal of M. aeruginosa was studied under different PS doses (Fig. 3). The removal of OD₆₈₁ in 251 252 death phase kept high efficiencies (>92.8%) for all the initial concentrations of PS 253 ranging from 0 mg/L to 1000 mg/L over an oxidation period of 120 min. However, for 254 the cells in log phase, the removal of OD₆₈₁ only increased from 21.8% to 83.7% with 255 the initial dosage of PS increasing from 0 mg/L to 1500 mg/L after reaction for 120 256 min, which was much lower than the removal efficiencies for the algal cells in death 257 phase. Similarly, the removal efficiency of Chl-a ranging from 90.7% to 98.5% in death phase by using UV/PS was also higher than that ranging from 87.9% to 97.6% 258 259 in log phase.

The easier removal of *M. aeruginosa* in death phase than that in log phase by using UV/PS process was supposed to be connected to the disparity of biological activity, EOM component, and morphologies of algal cells in different growth phases [56].

Besides, the autolysis of cells largely occurred in death phase [19], which might alsocontribute to the high removal efficiency.

265

266 3.4. Mineralization of AOM

267 Fig. 4 shows the residual TOC and extracellular DOC of *M. aeruginosa* suspension at 268 different PS doses. Negligible changes of TOC and extracellular DOC were observed 269 after single UV irradiation over a period of 120 min, indicating that limited 270 UV-induced mineralization of AOM occurred, which was in accordance with the 271 previous report [57]. On the other hand, the results reveal that the UV irradiation only had limited damage on the structure of algal cell and did not cause large-scale 272 cytoclasis [58], since an increase in DOC concentration would be found due to the 273 274 release of IOM from damaged algal cells [25, 48].

275 With the dosage of PS increasing from 0 mg/L to 1500 mg/L (about 6 mM), the 276 TOC and extracellular DOC were decreased from 12.50 and 7.59 mg/L to 1.63 and 277 1.61 mg/L, respectively, over a reaction period of 120 min. The significant removal of 278 extracellular DOC (decreased by 79.6% at 1500 mg/L PS dosage) indicates that UV/PS process can effectively mineralize AOM in the reaction solution. Moreover, 279 280 the result of higher reduced concentration of TOC than extracellular DOC in the 281 presence of PS can be explained by the fact that released dissolved IOM could make 282 up parts of mineralized DOC, reflecting that UV/PS oxidation can induce destruction 283 of algal cell structures. This speculation was supported by the result that residual TOC

(1.63 mg/L) and extracellular DOC (1.61 mg/L) were extremely approximate at an
initial PS dosage of 1500 mg/L.

The UV-Vis spectra of *M. aeruginosa* solution before and after UV/PS treatment were also studied (insert graph in Fig. 4). The absorbance of solution ranging from 200 to 700 nm was significantly decreased after UV/PS treatment, which was in accordance with the changes of TOC and DOC. It is noteworthy that, the peak at 681 nm was disappeared after treatment, suggesting the effective removal of algal cells in UV/PS system.

292

293 3.5. Evolution in fluorescence EEM spectra of extracellular AOM

294 The effects of UV/PS process treatment on fluorescence EEM spectra of extracellular 295 AOM are shown in Fig. 5 and Table S1. There were four fluorescence peaks at Ex/Em wavelengths of 270/442 nm (peak A), 350/432 nm (peak C), 230/330 nm (peak T₂), 296 297 and 280/330 nm (peak T1) in the control sample (Fig. 5a), which represented 298 fulvic-like, humic-like and protein-like substances, and dissolved microbial 299 metabolites, respectively [59, 60]. The strong signal at peak T₁ suggested high 300 concentration of protein-like substances contained in EOM, which was similar to 301 other reports [61, 62].

However, after UV/PS process treatment for 5 min, peaks A and C were observed to be significantly enhanced, accompanied with the disappearance of peaks T_1 and T_2 in contrast (Fig. 5b). This can be explained that UV/PS process was supposed to cause 305 cell secretion and decomposition which was expected to produce fulvic-like and306 humic-like substances [63, 64].

When the treatment time was over 30 min, all the four peaks were disappeared (Figs. 5c and 5d), whose possible reasons were proposed as below: 1) UV/PS process could further oxidize and even mineralize the formed humic-like and fulvic-like substances (Fig. 4); and 2) the large-scale cell cytoclasis and metabolites release mainly occurred in the initial reaction period under UV/PS oxidation, which was similar to the ozonation of *cyanobacteria* [65].

313 Additionally, comparative experiments of single UV irradiation, PS alone and 314 UV/PS process treatment on the changes of the fluorescence EEM spectra of extracellular AOM were conducted (Fig. S7). After treatment for 30 min, UV 315 316 irradiation caused significant decrease of peak T₁, but slight enhancement of peak C. 317 While the treatment of algae by PS alone had negligible impact on the fluorescence 318 EEM spectra of the extracellular AOM. In comparison with UV irradiation or PS 319 oxidation alone, UV/PS treatment caused the most significant decline of all the peaks 320 on the fluorescence EEM spectra. Thus, it is reasonable to conclude that the formed 321 reactive species, such as hydroxyl and sulfate radicals, in UV/PS process contribute to 322 the degradation of organic matter.

323

324 **3.6.** Cell integrity and cellular morphology

325 Fig. 6 shows the impacts of PS oxidation, UV irradiation, and UV/PS treatment on

326 cell integrity by using a flow cytometer to measure changes in cell permeability and 327 chlorophyll auto-fluorescence. P1 and P2 regions represent damaged and integrated 328 cell population, respectively. In Fig. 6a, the proportion of live cells was 96.1%, 329 reflecting that the vast majority of *M. aeruginosa* cells without any treatment were 330 intact. Compared with the control sample, the damaged cells shown in P1 region was 331 found to be only slightly increased by 1.8% after 2 days reaction with PS alone (Fig. 332 6b), indicating that individual PS oxidation played little to no roles in destruction of 333 *M. aeruginosa* cells . In contrast, UV was observed to be more effective in decreasing 334 the live algal cells as shown in regional P2 of Fig. 6c that the amount of live algal 335 cells was reduced. However, the amount of algal cells shown in regional P1 was also 336 decreased. Since the green fluorescence in channel FL1 (Regional P1) was from the 337 stained nucleic acid by SYTOX [43], the results can be explained by the fact that UV 338 irradiation could damage nucleic acid [66, 67]. Among the four treatment processes, 339 UV/PS process had the strongest ability to damage algal cells (Fig. 6d). After 340 treatment for 30 min, the integrated cell population was significantly reduced, 341 indicating that the reactive substances in UV/PS system, such as SO4⁻⁻ and HO⁺, can 342 strongly damage the cell integrity. No signals were found in regional P1 suggested 343 that UV/PS could further oxidize nucleic acid effectively due to the synergetic effect 344 of UV irradiation and the generated reactive radicals.

345 The surface morphologies of *M. aeruginosa* cells before and after UV/PS treatment 346 were observed using the SEM. Compared to the algal cells without any treatments

(Fig. 7a), the surface morphologies of cells were deformed and the cellular structure
was completely lost after the UV/PS treatment (Fig. 7b). Some apparent materials
were found to be released from the cells, which were probable to be amino acids and
carbohydrates [28].

351

352 **3.7.** Persulfate decomposition

353 Fig. 8a shows the persulfate decay under different initial PS doses ranging from 500 354 to 1500 mg/L. In each case, residual persulfate was less than 2% over a reaction 355 period of 120 min. The decomposition of persulfate followed the first-order kinetics model (Fig. 8b), suggesting that the presence of algal cells and AOM had little impact 356 357 on the self-decomposition of persulfate, which was supported by the insignificant 358 oxidation potential of PS on algal cells and AOM compared to UV/PS process (Figs. 1 359 and S7). The added PS was almost completely decomposed (\geq 99.9%) during the 360 reaction when the initial concentration of PS was less than 1000 mg/L. Even at initial 361 dosages of 1250 mg/L and 1500 mg/L, the average residual PS concentrations were 362 only 2.14 mg/L (about 0.2%) and 23.6 mg/L (about 1.6%) after treatment for 120 min, respectively. Thus, it is concluded that application of UV/PS process in the removal of 363 364 algae in natural water had little potential to cause undesired risk from the residual PS. 365

366 4. Conclusions

367 UV/PS process is demonstrated to be an effective technology for the removal of M.

368 aeruginosa in laboratory. The formed reactive species including SO4⁻⁻ and HO⁻ were proven to be the most important reasons for the removal of algal cells. Increasing PS 369 370 dosage could improve the removal of algal cells and the mineralization of AOM. With 371 the addition of 1500 mg/L (about 6 mM) PS, removal rates up to 98.2% and 99.8% of 372 algal cells and Chl-a, respectively, were achieved in UV/PS system over a reaction 373 period of 120 min. M. aeruginosa in death phase was easier to be removed by using 374 UV/PS treatment than that in log phase. Although the UV/PS process significantly 375 damaged the structures of algal cell and caused the release of IOM, these derived 376 organic compounds along with the dissolved EOM could be further mineralized in the 377 subsequent reaction period. The fast decomposition of persulfate might insure the safety of using UV/PS to remove algal cells to some extent. 378

379

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384

385 Appendix A. Supplementary data

386 Supplementary data (Figs. S1 to S7 and Tab. S1) associated with this article can be

387 found, in the online version.

389 **References**

- 390 [1] B.E. Lapointe, L.W. Herren, D.D. Debortoli, M.A. Vogel, Evidence of
 391 sewage-driven eutrophication and harmful algal blooms in Florida's Indian River
 392 Lagoon, Harmful Algae 43 (2015) 82-102.
- 393 [2] J.D. Isaacs, W.K. Strangman, A.E. Barbera, M.A. Mallin, M.R. McIver, J.L.C.
- 394 Wright, Microcystins and two new micropeptin cyanopeptides produced by
- 395 unprecedented Microcystis aeruginosa blooms in North Carolina's Cape Fear River,
- 396 Harmful Algae 31 (2014) 82-86.
- 397 [3] M.A. Burford, S.A. Johnson, A.J. Cook, T.V. Packer, B.M. Taylor, E.R. Townsley,
- 398 Correlations between watershed and reservoir characteristics, and algal blooms in 399 subtropical reservoirs, Water Research 41 (2007) 4105-4114.
- 400 [4] V.H. Smith, Eutrophication of freshwater and coastal marine ecosystems: a global
- 401 problem, Environmental Science and Pollution Research 10 (2003) 126.
- 402 [5] R.K. Henderson, A. Baker, S.A. Parsons, B. Jefferson, Characterisation of
- 403 algogenic organic matter extracted from cyanobacteria, green algae and diatoms,
- 404 Water Research 42 (2008) 3435-3445.
- 405 [6] J.O. Jo, S.D. Kim, H.J. Lee, Y.S. Mok, Decomposition of taste-and-odor
- 406 compounds produced by cyanobacteria algae using atmospheric pressure plasma
- 407 created inside a porous hydrophobic ceramic tube, Chemical Engineering Journal 247
- 408 (2014) 291-301.
- 409 [7] J.P. Berry, P.D. Gibbs, M.C. Schmale, M.L. Saker, Toxicity of cylindrospermopsin,

and other apparent metabolites from Cylindrospermopsis raciborskii and
Aphanizomenon ovalisporum, to the zebrafish (Danio rerio) embryo, Toxicon :
official journal of the International Society on Toxinology 53 (2009) 289-299.

- 413 [8] R. Bhattacharya, K. Sugendran, R.S. Dangi, P.V.L. Rao, Toxicity evaluation of
- 414 freshwater cyanobacterium Microcystis aeruginosa PCC 7806: II Nephrotoxicity in
- 415 rats, Biomedical and Environmental Sciences 10 (1997) 93-101.
- 416 [9] C. MacKintosh, K.A. Beattie, S. Klumpp, P. Cohen, G.A. Codd, Cyanobacterial
- 417 microcystin-LR is a potent and specific inhibitor of protein phosphatases 1 and 2A
- 418 from both mammals and higher plants, FEBS letters 264 (1990) 187-192.
- 419 [10] M.J. Bakke, T.E. Horsberg, Effects of algal-produced neurotoxins on metabolic
- 420 activity in telencephalon, optic tectum and cerebellum of Atlantic salmon (Salmo
- 421 salar), Aquatic toxicology 85 (2007) 96-103.
- 422 [11] J. Fang, J. Ma, X. Yang, C. Shang, Formation of carbonaceous and nitrogenous
- 423 disinfection by-products from the chlorination of Microcystis aeruginosa, Water424 Research 44 (2010) 1934-1940.
- 425 [12] Y.S. Lui, J.W. Qiu, Y.L. Zhang, M.H. Wong, Y. Liang, Algal-derived organic
- 426 matter as precursors of disinfection by-products and mutagens upon chlorination,
- 427 Water Research 45 (2011) 1454-1462.
- 428 [13] X. Yang, W. Guo, Q. Shen, Formation of disinfection byproducts from
 429 chlor(am)ination of algal organic matter, Journal of hazardous materials 197 (2011)
 430 378-388.

- 431 [14] J. Edzwald, Algae, bubbles, coagulants, and dissolved air flotation, Water
 432 Science & Technology 27 (1993) 67-81.
- 433 [15] C.W. Chow, M. Drikas, J. House, M.D. Burch, R.M. Velzeboer, The impact of
- 434 conventional water treatment processes on cells of the cyanobacterium Microcystis
- 435 aeruginosa, Water Research 33 (1999) 3253-3262.
- 436 [16] M.R. Teixeira, M.J. Rosa, Comparing dissolved air flotation and conventional
- 437 sedimentation to remove cyanobacterial cells of Microcystis aeruginosa, Separation
- 438 and Purification Technology 52 (2006) 84-94.
- 439 [17] Q. Shen, J. Zhu, L. Cheng, J. Zhang, Z. Zhang, X. Xu, Enhanced algae removal
- 440 by drinking water treatment of chlorination coupled with coagulation, Desalination441 271 (2011) 236-240.
- 442 [18] D.L. Widrig, K.A. Gray, K.S. McAuliffe, Removal of algal-derived organic
- 443 material by preozonation and coagulation: Monitoring changes in organic quality by
- 444 pyrolysis-GC-MS, Water Research 30 (1996) 2621- 2632.
- 445 [19] M. Pivokonsky, O. Kloucek, L. Pivokonska, Evaluation of the production,
- 446 composition and aluminum and iron complexation of algogenic organic matter, Water447 Research 40 (2006) 3045-3052.
- 448 [20] M. Ma, R. Liu, H. Liu, J. Qu, W. Jefferson, Effects and mechanisms of
- 449 pre-chlorination on Microcystis aeruginosa removal by alum coagulation:
- 450 Significance of the released intracellular organic matter, Separation and Purification
- 451 Technology 86 (2012) 19-25.

- 452 [21] H. Hori, A. Yamamoto, E. Hayakawa, S. Taniyasu, N. Yamashita, S. Kutsuna, H.
- 453 kiatagawa, R. Arakawa, Efficient decomposition of environmentally persistent
 454 perfluorocarboxylic acids by use of persulfate as a photochemical oxidant,
 455 Environmental Science & Technology 39 (2005) 2383-2388.
- 456 [22] A. Tsitonaki, B.F. Smets, P.L. Bjerg, Effects of heat-activated persulfate oxidation
- 457 on soil microorganisms, Water Research 42 (2008) 1013-1022.
- 458 [23] H. Liu, T.A. Bruton, F.M. Doyle, D.L. Sedlak, In Situ Chemical Oxidation of
- 459 Contaminated Groundwater by Persulfate: Decomposition by Fe(III)- and
- 460 Mn(IV)-Containing Oxides and Aquifer Materials, Environmental Science &
- 461 Technology 48 (2014) 10330-10336.
- 462 [24] S. Su, W. Guo, C. Yi, Y. Leng, Z. Ma, Degradation of amoxicillin in aqueous
 463 solution using sulphate radicals under ultrasound irradiation, Ultrason Sonochem 19

464 (2012) 469-474.

- 465 [25] Y. Ji, Y. Fan, K. Liu, D. Kong, J. Lu, Thermo activated persulfate oxidation of
 466 antibiotic sulfamethoxazole and structurally related compounds, Water Research 87
- 467 (2015) 1-9.
- 468 [26] T.K. Lau, W. Chu, N.J. Graham, The aqueous degradation of butylated
- 469 hydroxyanisole by $UV/S_2O_8^2$: study of reaction mechanisms via dimerization and
- 470 mineralization, Environmental Science & Technology 41 (2007) 613-619.
- 471 [27] P. Neta, V. Madhavan, H. Zemel, R.W. Fessenden, Rate constants and mechanism
- 472 of reaction of sulfate radical anion with aromatic compounds, Journal of the American

- 473 Chemical Society 99 (1977) 163-164.
- 474 [28] S. Ahn, T.D. Peterson, J. Righter, D.M. Miles, P.G. Tratnyek, Disinfection of
 475 ballast water with iron activated persulfate, Environmental Science & Technology 47
 476 (2013) 11717-11725.
- 477 [29] G. Mark, M.N. Schuchmann, H.P. Schuchmann, C. von Sonntag, The photolysis
- 478 of potassium peroxodisulphate in aqueous solution in the presence of tert-butanol: a
- 479 simple actinometer for 254 nm radiation, Journal of Photochemistry and Photobiology
- 480 A: Chemistry 55 (1990) 157-168.
- 481 [30] P. Neta, R.E. Huie, A.B. Ross, Rate constants for reactions of inorganic radicals
- 482 in aqueous solution, Journal of Physical and Chemical Reference Data 17 (1988)483 1027-1284.
- 484 [31] Y. Tao, X. Zhang, D.W. Au, X. Mao, K. Yuan, The effects of sub-lethal UV-C
 485 irradiation on growth and cell integrity of cyanobacteria and green algae,
 486 Chemosphere 78 (2010) 541-547.
- 487 [32] H. Ou, N. Gao, Y. Deng, H. Wang, H. Zhang, Inactivation and degradation of
- 488 Microcystis aeruginosa by UV-C irradiation, Chemosphere 85 (2011) 1192-1198.
- 489 [33] M.G. Antoniou, A. Armah, D.D. Dionysiou, Degradation of microcystin-LR
- 490 using sulfate radicals generated through photolysis, thermolysis and e- transfer
- 491 mechanisms, Applied Catalysis B: Environmental 96 (2010) 290-298.
- 492 [34] N.S. Shah, X. He, H.M. Khan, J.A. Khan, K.E. O'Shea, D.L. Boccelli, D.D.
- 493 Dionysiou, Efficient removal of endosulfan from aqueous solution by UV-C/peroxides:

- 494 a comparative study, Journal of hazardous materials 263 (2013) 584-592.
- 495 [35] V.K. Sharma, T.M. Triantis, M.G. Antoniou, X. He, M. Pelaez, C. Han, W. Song,
- 496 K.E. O'Shea, A. Armah, T. Kaloudis, Destruction of microcystins by conventional and
- 497 advanced oxidation processes: a review, Separation and Purification Technology 91
- 498 (2012) 3-17.
- 499 [36] P. Xie, J. Ma, J. Fang, Y. Guan, S. Yue, X. Li, L. Chen, Comparison of
- 500 Permanganate Preoxidation and Preozonation on Algae Containing Water: Cell
- 501 Integrity, Characteristics, and Chlorinated Disinfection Byproduct Formation,
- 502 Environmental Science & Technology 47 (2013) 14051-14061.
- 503 [37] R. Rippka, J. Deruelles, J.B. Waterbury, M. Herdman, R.Y. Stanier, Generic
- 504 assignments, strain histories and properties of pure cultures of cyanobacteria, Journal
- 505 of General microbiology 111 (1979) 1-61.
- 506 [38] P. Li, Y. Song, S. Yu, Removal of Microcystis aeruginosa using hydrodynamic
- 507 cavitation: Performance and mechanisms, Water Research 62 (2014) 241-248.
- 508 [39] R.O. Rahn, Potassium iodide as a chemical actinometer for 254 nm radiation: use
- of lodate as an electron scavenger, Photochemistry and Photobiology 66 (1997)450-455.
- 511 [40] W. Liang, J. Qu, L. Chen, H. Liu, P. Lei, Inactivation of Microcystis aeruginosa
- 512 by continuous electrochemical cycling process in tube using Ti/RuO2 electrodes,
- 513 Environmental Science & Technology 39 (2005) 4633-4639.
- 514 [41] G. Zhang, P. Zhang, B. Wang, H. Liu, Ultrasonic frequency effects on the

- 515 removal of Microcystis aeruginosa, Ultrasonics Sonochemistry 13 (2006) 446-450.
- 516 [42] O. Holm-Hansen, B. Riemann, Chlorophyll a determination: improvements in
- 517 methodology, Oikos (1978) 438-447.
- 518 [43] R.I. Daly, L. Ho, J.D. Brookes, Effect of chlorination on Microcystis aeruginosa
- 519 cell integrity and subsequent microcystin release and degradation, Environmental
- 520 Science & Technology 41 (2007) 4447-4453.
- 521 [44] I. Kolthoff, E. Carr, Volumetric determination of persulfate in presence of organic
- 522 substances, Analytical Chemistry 25 (1953) 298-301.
- 523 [45] H. Ou, N. Gao, Y. Deng, J. Qiao, H. Wang, Immediate and long-term impacts of
- 524 UV-C irradiation on photosynthetic capacity, survival and microcystin-LR release risk
- of Microcystis aeruginosa, Water Research 46 (2012) 1241-1250.
- 526 [46] P. Xie, J. Ma, W. Liu, J. Zou, S. Yue, X. Li, M.R. Wiesner, J. Fang, Removal of
- 527 2-MIB and geosmin using UV/persulfate: Contributions of hydroxyl and sulfate
- 528 radicals, Water research 69 (2015) 223-233.
- 529 [47] X.-Y. Yu, Z.-C. Bao, J.R. Barker, Free Radical Reactions Involving Cl, Cl2-, and
- 530 SO4-in the 248 nm Photolysis of Aqueous Solutions Containing S2O82-and Cl, The
- 531 Journal of Physical Chemistry A 108 (2004) 295-308.
- 532 [48] G.P. Anipsitakis, D.D. Dionysiou, M.A. Gonzalez, Cobalt-mediated activation of
- 533 peroxymonosulfate and sulfate radical attack on phenolic compounds. Implications of
- chloride ions, Environmental science & technology 40 (2006) 1000-1007.
- 535 [49] H.V. Lutze, N. Kerlin, T.C. Schmidt, Sulfate radical-based water treatment in

- 536 presence of chloride: Formation of chlorate, inter-conversion of sulfate radicals into
- 537 hydroxyl radicals and influence of bicarbonate, Water research 72 (2015) 349-360.
- 538 [50] Y. Tao, X. Mao, J. Hu, H.O. Mok, L. Wang, D.W. Au, J. Zhu, X. Zhang,
- 539 Mechanisms of photosynthetic inactivation on growth suppression of Microcystis
- 540 aeruginosa under UV-C stress, Chemosphere 93 (2013) 637-644.
- 541 [51] H. Liu, T.A. Bruton, W. Li, J. Van Buren, C. Prasse, F.M. Doyle, D.L. Sedlak,
- 542 Oxidation of Benzene by Persulfate in the Presence of Fe (III)-and Mn
- 543 (IV)-Containing Oxides: Stoichiometric Efficiency and Transformation Products,
- 544 Environmental science & technology (2015).
- 545 [52] R. Szabó, C. Megyeri, E. Illés, K. Gajda-Schrantz, P. Mazellier, A. Dombi,
- 546 Phototransformation of ibuprofen and ketoprofen in aqueous solutions, Chemosphere

547 84 (2011) 1658-1663.

- 548 [53] B. Lee, M. Lee, Decomposition of 2, 4, 6-trinitrotoluene (TNT) by gamma
- 549 irradiation, Environmental science & technology 39 (2005) 9278-9285.
- 550 [54] D. Guimaraes, V.A. Leao, Batch and fixed-bed assessment of sulphate removal
- by the weak base ion exchange resin Amberlyst A21, Journal of hazardous materials280 (2014) 209-215.
- 553 [55] M. Melkonian, Release of Dissolved Organic Substances by the Green Alga
- 554 Fritschiella tuberosalyeng. (Chaetophorineae) During Different Growth Phases,
- 555 Zeitschrift für Pflanzenphysiologie 94 (1979) 125-133.
- 556 [56] J. Huang, N. Graham, M.R. Templeton, Y. Zhang, C. Collins, M. Nieuwenhuijsen,

- A comparison of the role of two blue-green algae in THM and HAA formation, Water
 Research 43 (2009) 3009-3018.
- 559 [57] H. Xu, H. Jiang, UV-induced photochemical heterogeneity of dissolved and
- 560 attached organic matter associated with cyanobacterial blooms in a eutrophic
- 561 freshwater lake, Water research 47 (2013) 6506-6515.
- 562 [58] H. Sakai, H. Katayama, K. Oguma, S. Ohgaki, Kinetics of Microcystis
 563 aeruginosa Growth and Intracellular Microcystins Release after UV Irradiation,
 564 Environmental Science & Technology 43 (2009) 896-901.
- 565 [59] N. Hudson, A. Baker, D. Reynolds, Fluorescence analysis of dissolved organic
- 566 matter in natural, waste and polluted waters A review, River Research and 567 Applications 23 (2007) 631-649.
- 568 [60] P.G. Coble, Characterization of marine and terrestrial DOM in seawater using
- 569 excitation. Emission matrix spectroscopy, Marine Chemistry 51 (1996) 325-346.
- 570 [61] W. Huang, H. Chu, B. Dong, J. Liu, Evaluation of different algogenic organic
- 571 matters on the fouling of microfiltration membranes, Desalination 344 (2014)572 329-338.
- 573 [62] H. Ou, N. Gao, Y. Deng, J. Qiao, K. Zhang, T. Li, L. Dong, Mechanistic studies
- 574 of Microcystic aeruginosa inactivation and degradation by UV-C irradiation and
- 575 chlorination with poly-synchronous analyses, Desalination 272 (2011) 107-119.
- 576 [63] E.J. Rochelle-Newall, T.R. Fisher, Production of chromophoric dissolved organic
- 577 matter fluorescence in marine and estuarine environments: an investigation into the

- 578 role of phytoplankton, Marine Chemistry 77 (2002) 7-21.
- 579 [64] L. Li, C. Shao, T.-F. Lin, J. Shen, S. Yu, R. Shang, D. Yin, K. Zhang, N. Gao,
- 580 Kinetics of cell inactivation, toxin release, and degradation during permanganation of
- 581 Microcystis aeruginosa, Environmental science & technology 48 (2014) 2885-2892.
- 582 [65] H. Miao, W. Tao, The mechanisms of ozonation on cyanobacteria and its toxins
- removal, Separation and Purification Technology 66 (2009) 187-193.
- 584 [66] R.P. Sinha, D.-P. Häder, UV-induced DNA damage and repair: a review,
- 585 Photochemical & Photobiological Sciences 1 (2002) 225-236.
- 586 [67] J. Cadet, C. Anselmino, T. Douki, L. Voituriez, New trends in photobiology:
- 587 Photochemistry of nucleic acids in cells, Journal of Photochemistry and Photobiology
- 588 B: Biology 15 (1992) 277-298.